

RESEARCH ARTICLE

Going Cheap: Determinants of Bird Price in the Taiwanese Pet Market

Shan Su^{1,2,3,*}, Phillip Cassey³, Miquel Vall-Ilosera³, Tim M. Blackburn^{2,3,4}

1 Institute of Zoology, Zoological Society of London, Regent's Park, London, United Kingdom, **2** Research Department of Genetics, Evolution and Environment, University College London, Gower Street, London, United Kingdom, **3** School of Biological Sciences, University of Adelaide, Adelaide, Australia, **4** Centre for Invasion Biology, Department of Botany and Zoology, Stellenbosch University, Stellenbosch, South Africa

☞ These authors contributed equally to this work.

* su.shan@ioz.ac.uk



Abstract

Background

International wildlife trade is the largest emerging source of vertebrate invasive alien species. In order to prevent invasions, it is essential to understand the mechanics of trade and, in particular, which traded species are most likely to be released or escape into the wild. A species' economic value is a key factor, because we expect cheaper species to be less assiduously secured against escaping, and more likely to be deliberately released. Here, we investigate determinants of the price of species in the Taiwanese bird trade. Taiwan is an international hub for bird trade, and several native species are threatened by alien bird species.

Methodology

We investigated the relationship between the traded species sale price in Taiwan and the species availability for trade (the number of birds for sale, geographic range size and their origin, conservation and CITES status) and traits (body size, coloration, song attractiveness). We used phylogenetic generalized least squares models, with multi-model inference, to assess the variables that are best related to the price of birds in the Taiwanese pet trade.

Principal Findings / Conclusions

We found that species available for sale in larger numbers, native to Taiwan, not globally endangered, and small-bodied are all relatively cheaper, as too are species lacking yellow coloration and without attractive songs. Our models of price revealed high levels of phylogenetic correlation, and hence that closely related species tended to be sold for similar prices. We suggest that, on the basis of price, native species are more likely to be deliberately or accidentally released than alien species. Nevertheless, our survey of bird shops recorded 160 species alien to Taiwan (7,631 individuals), several of which are for sale cheaply and in large numbers. Alien bird species in trade therefore present an ongoing, non-trivial invasion risk on the island.

OPEN ACCESS

Citation: Su S, Cassey P, Vall-Ilosera M, Blackburn TM (2015) Going Cheap: Determinants of Bird Price in the Taiwanese Pet Market. PLoS ONE 10(5): e0127482. doi:10.1371/journal.pone.0127482

Academic Editor: Sergio A Lambertucci, INIBIOMA (Universidad Nacional del Comahue-CONICET), ARGENTINA

Received: December 29, 2014

Accepted: April 14, 2015

Published: May 27, 2015

Copyright: © 2015 Su et al. This is an open access article distributed under the terms of the [Creative Commons Attribution License](https://creativecommons.org/licenses/by/4.0/), which permits unrestricted use, distribution, and reproduction in any medium, provided the original author and source are credited.

Data Availability Statement: All relevant data are within the paper and its Supporting Information files.

Funding: PC is an ARC Future Fellow (FT0991420). This study received financial support from Institute of Zoology, Zoological Society of London (zsl.org), and from the Invasive Animals CRC (to MV). The funders had no role in study design, data collection and analysis, decision to publish, or preparation of the manuscript.

Competing Interests: The authors have declared that no competing interests exist.

Introduction

Geographical barriers that have naturally set limits species distributions are increasingly being breached as species are moved around by international trade [1–4]. This is a cause for considerable concern because alien species can have substantial negative impacts on the recipient regions [5–7], including causing the extinction of native species [6,8] and the homogenization of ecological assemblages (with a concomitant decrease in species diversity) [9].

One of the major pathways of human-mediated movement is the trade in wild animals and plants, including the trade in cage birds [10]. Captivity places constraints on the invasion process [4,11], because traded species are not directly introduced after transport and not all species in trade are introduced. Nevertheless, the wildlife vertebrate trade is an increasing source of invasive species [12–14]. Alien species in trade have been shown to be correlated, in terms of identity [15] and abundance [16], with those species that subsequently establish alien populations. Indeed, the most important correlate of variation in the number of invasive alien species across countries is currently the volume of merchandise imports [17], suggesting that international trade is now the primary driver of species invasions (see also Colunga-Garcia *et al.* [18]). To characterise invasions effectively, it is therefore important to understand the mechanics of the wildlife trade, and in particular, which traded species are most likely to be released or escape into the wild.

The amount of care devoted to keeping a bird is likely to be associated with its value [19]. Therefore, we have assumed that sale price is one of the factors that will influence the likelihood that traded species are introduced into the wild. More valuable species are expected to receive better care and hence to be more assiduously secured against escaping, and less likely to be deliberately released by their owners. Here, we use the example of the cage bird trade in Taiwan to explore reasons for variation in the price of species for sale.

Taiwan presents an excellent example for studying features of the bird trade in Asian countries [20]. First, cultural practices relating to the bird trade, such as prayer release, singing competitions and ‘bird-walking’ (the avian equivalent of dog-walking, where birds are taken out in cages for fresh air), are very popular in Taiwan. Prayer animal release is a common religious activity in Asia [20]: it is estimated that more than 200 million animals are set free through this practice in Taiwan every year [21,22]. Such a high frequency of releases may be an important factor contributing to invasions by alien bird species [5,23–26]. Second, Taiwan is a major wildlife trade hub from where species are re-exported to other Asian countries, such as Malaysia and Singapore [20,27,28]. Third, Taiwan is separated from the nearest continent by a >100 km ocean strait, meaning that it is relatively straightforward to define native versus alien species. Finally, alien bird species have proved to be a conservation threat in Taiwan through actual and potential hybridization with native bird species [29–31].

We predicted that the price of traded birds in the Taiwanese pet market would largely be determined by a species’ availability, because supply costs would be lower for common species, and because people will pay higher prices for rarer species (the anthropogenic Allee effect [32]). We therefore hypothesised a negative relationship between price and the number of individuals of a species offered for sale. Species that are more widespread are normally more abundant [33,34], and more available for trade [24]. We hypothesised that species that have more extensive native distributions would be less expensive. Similarly, we hypothesised that native species and species coming from closer regions, species not considered as threatened on the IUCN Red List, or species not listed on CITES Appendices, would also be cheaper. However, species availability is not the only factor suggested to affect market value. An extensive body of literature has shown that particular species traits can affect human attitudes towards species, including willingness to pay. Animal body size is a trait that influences human preference [35],

and larger-sized species are more valued [36]. Body size is negatively related to bird species richness [37,38] and abundance [34,39], both of which may affect the availability of large-bodied birds for sale. Large-bodied birds also tend to be longer-lived and reproduce more slowly, and so will take longer to rear and be less productive if captive bred. For all these reasons, we therefore predicted that species with small body sizes should be less expensive. We also expected that more colourful species would be more attractive as cage birds [40–42], and hence predicted that drabber species would sell for lower prices]. Finally, species are traded for singing competitions in Asia, and so we hypothesised that species with less attractive songs would fetch a lower price.

Methods

Data

To obtain data on the occurrence and price of bird species in Taiwanese pet shops, we (SS) conducted non-structured interviews between August and November 2012. Data on the locations of pet shops were obtained from the Commerce Industrial Services Portal, telephone directories, previous surveys and pet shop-related blogs. In this way, we located a total of 154 pet shops. This excludes game pigeon stores, pet product-only and aquarium shops. During the survey, we found that some of the shops had closed down, while some owners declined to respond to the survey. Therefore a total of 72 shops were surveyed in seven cities. Of the final survey sites, 32 were in Taipei City (and New Taipei City), seven in Hsinchu City, 11 in Taichung City, one in Taitung City, 14 in Kaohsiung City, five in Pingtung County and two shops in Hualien City. All shops were visited at least once with the aim of obtaining data on three variables: (i) the identity of species displayed for sale, (ii) their price (in Taiwanese Dollars, or TWD) and (iii) the numbers of birds of each species displayed. During each visit, SS first enquired about the identity of the species available for sale, their price and then counted and recorded the number of birds of each species displayed.

Ethics approval for the study was given by the ZSL Ethics Committee. The non-structured interviews of the shop owners did not involve questions about their personal information or identity. This is because the sources of birds in trade in Taiwan are frequently illegal, and therefore owners will not provide signed documents and/or written consent. We aimed to record information on the numbers of species, their identity and their price in the shops, and we obtained shop owners' verbal agreements to obtain such information. The consent procedure was documented in writing by SS and had been approved by the Ethics Committee.

We identified all taxa to species-level except for three white-eye species (Oriental White-eye *Zosterops palpebrosus*, Lowland White-eye *Z. meyeri* and Japanese White-eye *Z. japonicus*). The three White-eye species could be identified only by the owners, but the names they give can be unreliable as the sources of some of the White-eye species can be illegal. Two Myna species (Javan Myna *Acridotheres javanicus* and Great Myna *A. grandis*) were identified to species level, but these species are frequently stocked in mixed species cages with more than 50 birds in each cage. This makes it difficult to count the number of each species in a cage, and we therefore assumed that all mynas in any given cage belonged to the majority species. Sometimes the cages are packed and stacked on top of one-another, which makes species-level identification difficult. In this case, we assumed that all mynas in stacked cages belonged to the majority species from the visible cages. Where owners or staff were not prepared to reveal species identities, we identified birds to species level using monographs and bird guides, often from photographs taken by SS. In this way, a total of 247 bird species were identified. We used only species recorded in the first visit to each shop for analysis ($n = 221$) to standardise estimates of numbers for sale per shop. Furthermore, the only two Nearctic species recorded for sale (Northern

Mockingbird *Mimus polyglottos* and Red-bellied Woodpecker *Melanerpes carolinus*) were removed to ease regional comparisons. One further species (Taiwan Hwamei *Garrulax taewanus*) was excluded because an estimate of price was not available. For PGLS analysis (see below) we excluded Red-collared Lorikeet *Trichoglossus rubritorquis* when phylogenetic trees were selected based on [43], because the authors treat it as a subspecies of Rainbow Lorikeet *Trichoglossus haematodus* (resulting in $n = 217$).

We based our analysis on the global taxonomic list of 9,993 extant bird species from Jetz *et al.* [43]. The scientific names of the species followed Birdlife Version 3 and IOC Version 2.7 taxonomies, family names followed the Birdlife guide and the order names followed Sibley & Monroe (1990) [43]. We identified species as native or alien in Taiwan based on the CWBF Checklist of the Birds of Taiwan [44], adjusted by excluding transient migrants, pelagic seabirds and vagrants.

We collated information for each of the species on the following variables:

1. Geographical range size (km^2): a measure of native geographical range extent [45,46], acquired from the data applied in [47][47]. Maps of total native breeding area for all species were converted to equal area grid polygons with a cell size of $96.3 \text{ km} \times 96.3 \text{ km}$. This specified a scale identical to 1° grids at the equator. The breeding range of each species was estimated by summing the areas of the cells in which they occurred [47,48].
2. Region of origin: the biogeographic region from which the native populations of alien species originate, defined using the World Wildlife Fund eco-regions map [48,49]. Species were assigned to the region in which the largest part of their geographic range fell.
3. Body mass (g): a measure bird species body size from Dunning [50], augmented by data used in Olson *et al.* [51].
4. Conservation status of species: The conservation status of the identified species was obtained from the IUCN Red List of Threatened Species [52]. Species recorded in our survey were classified into the following categories, treated as a multi-level categorical variable: least concern (LC), near threatened (NT), vulnerable (VU) and endangered (EN). All categories were tested in our univariate analysis. We then separated species by whether they were in the threatened categories (NT, VU and EN) or not (LC) in our multivariate analysis. None of the species in our survey of pet shops was data deficient (DD), critically endangered (CR) or extinct in the wild (EW).
5. International trade status: The status of the identified species was obtained from Convention of International Trade in Endangered Species of Wild Fauna and Flora (CITES Appendices I, II and III). We separated species by whether they were included in Appendices I, II and III (scored 1) or not (scored 0).
6. Colour: Adult males of the nominal subspecies in breeding plumage (or females for species with reverse sexual dimorphism) were assessed for colouration on the basis of the plates in the Handbook of the Birds of the World. The colour of fourteen body parts of the bird (bill, face, cheek, head, throat, breast, belly, flank, back, wings, tail, rump, vent and legs) was assessed by eye, and the presence of eleven colours (blue, green, red, orange, brown, pink (including purple and violet), yellow, black, grey, white and ivory (pale flesh colour)) was scored. Two metrics were calculated from these data: 1) the percentage of the body surface covered by each colour, calculated over the fourteen body parts of the bird scored; and 2) colour diversity, calculated as the number of different colours on a given species (excluding colours covering $<3\%$ of the body).

7. Song attractiveness: We used the number of song tracks lodged for a given species on the Xeno-Canto website [www.xeno-canto.org] to derive a metric of the attractiveness of bird song to humans (following [53]). Xeno-Canto contains more than 115,000 independent recordings of bird songs, which cover about 80–85% of the extant species, with up to 335 recordings per species. Only two of the species in our data set do not have any song recordings on Xeno-Canto, while the range for the remainder is 1–162 records. We used the `glm` function in R v.3.0.3 [54] to model the log+1 number of songs lodged on Xeno-Canto as a function of log geographic range size. It has been shown that larger residuals are correlated with the characteristics of songs associated with attractiveness to people [53].

Of the nine predictor variables in the univariate analysis (see below), bird numbers, species status, region of origin, IUCN list and CITES listing, body mass and each colour, breeding range size and the attractiveness of bird song are available for 217 species. We log-transformed the data on body mass, number of birds of each species for sale, percentage of covered colour and their price, and applied arcsine square root transformations to the percentage of body surface covered in each colour.

Analysis

All analyses were conducted in R version 3.0.3 [54].

We constructed phylogenetic generalized least squares (PGLS) models for the prices of species recorded in Taiwanese pet shops (response variable) compared with all nine univariate predictor variables, using the *pgls* function in the package *caper* (region of origin, species status, IUCN status, and CITES list were treated as categorical variables). The *pgls* model specifies the structure of the variance—covariance matrix reflecting the phylogenetic associations among the species, and also includes an estimated parameter (λ) that governs the strength of phylogenetic signal in the dependent variable. Thus, *pgls* does not require the *a priori* assignment of phylogenetic signal but instead allows the data to dictate its magnitude in the statistical model [55]. Clearly, price is not an evolved trait, but it is correlated with species traits that are (see [Results](#)), and our phylogenetic approach provides a way to analyse the influence of these traits under a specified evolutionary hypothesis (i.e. a phylogeny) in a way that accounts for the non-independence of species in terms of price (c.f. [56]). The significance of colour in univariate analysis was assessed after a Bonferroni correction, given that we test for the effects of 11 different colours. Collinearity amongst the different colour variables was low. Ivory colouration was found to have the strongest correlation with other colour variables; it exhibits the strongest positive correlation to brown (Pearson's $r = 0.37$, $n = 217$, $P < 0.001$) and negative to black (Pearson's $r = -0.33$, $n = 217$, $P < 0.001$). To find the most likely model for bird price, we fitted a global PGLS model including all predictors. We then used the *dredge* and *model.avg* functions (package *MuMIn*) to fit all possible models from these predictor variables, to identify the most likely models, and to calculate Akaike weights and variable importance (the sum of the Akaike weights across all models including that variable) based on the Akaike Information Criterion corrected for small sample sizes (AICc).

We based our phylogenetic analyses on the phylogenetic tree for birds proposed by [43]. The structure of this phylogenetic tree is not known for certain, and so we incorporated uncertainty over the true phylogenetic relationship by repeating our analyses over a number of different phylogenetic trees for our species, sampled at random from www.birdtree.org [43]. Initially, we used a sample of 100 random trees to calculate our univariate relationships. We then used ten trees (the smaller sample due to constraints on computer time necessary for the multi-model approach over hundreds of possible models) to identify the most likely set of models (those with $\Delta\text{AICc} < 4$ of the best model) from our global PGLS model, and a further

random selection of 100 phylogenetic trees to assess the fit and importance of models in this set. Thus, the simulation produced median estimates (over these trees) of the parameters of the best-fit models incorporating phylogenetic uncertainty. This approach allowed us to provide estimates of the variance in regression coefficients, which we calculated as the 5th and 95th percentiles in the coefficients across the 100 runs.

Results

The total number of birds for sale in the pet shops in our survey across all 217 species was 26,165, of which 7,612 individuals were of species alien to Taiwan ($n = 159$). The maximum recorded number of birds per species was 7,420 for Red Turtle Dove *Streptopelia tranquebarica*, while the highest single bird count in a single shop was 4,000 individuals, for this same species. The ten most abundant species in Taiwanese pet shops are listed in [Table 1](#). These ten species comprised 80% of all individuals for sale, while the five native species in [Table 1](#) account for 70% of all birds for sale. More than half of the alien bird individuals for sale came from just five species ([Table 1](#)). 67 of the alien species (2,778 individuals) for sale in the pet shops were parrots.

The prices of the 217 identified species sold varied widely, from US\$0.86 (at an exchange rate of US\$1:TWD30) per individual for Eurasian Tree Sparrow *Passer montanus* to US\$8,000 for a single Golden Parakeet *Guaruba guarouba*. The total market value of displayed birds in the surveyed pet shops during the survey time was more than US\$0.8 million, of which more than US\$0.5 million (> 65%) was contributed by alien species.

Univariate PGLS models revealed that the prices of birds for sale in pet shops were related to the number of birds of each species for sale ([Fig 1](#)), body mass, whether or not the species is alien to Taiwan ([Fig 1](#)), whether or not the species is listed in CITES Appendices, breeding geographic range size, song attractiveness ([Table 2](#)) and colour (the percentage of a species' body that was yellow or grey; [Table 3](#)). There was also an effect of a species' categorisation on the IUCN Red List (adjusted $r^2 \pm \text{s.e.} = 0.084 \pm 0.0008$, $F_{3,177} \pm \text{s.e.} = 5.41 \pm 0.059$, $p \pm \text{s.e.} = 0.002 \pm 0.0001$, $\lambda \pm \text{s.e.} = 0.952 \pm 0.0007$), and a marginal effect of biogeographic realm of origin (adjusted $r^2 \pm \text{s.e.} = 0.091 \pm 0.0007$, $F_{4,88} \pm \text{s.e.} = 2.21 \pm 0.01$, $p \pm \text{s.e.} = 0.076 \pm 0.002$, $\lambda \pm \text{s.e.} = 0.95 \pm 0.0004$): in general, less threatened species and species from realms closer to Taiwan (Palearctic, which include Taiwanese native species) were cheaper.

Model selection on the global model for the price of birds in Taiwanese pet shops identified 13 models for which the median ΔAICc across the ten trees was <4 relative to the most likely model ([Table 4](#)). As an indicator of fit, the variables in the most likely model explained 61% of the variance in bird price. The simulation over 100 randomly chosen but likely phylogenetic trees for each of these best 13 PGLS models recovered four variables—body mass, number of birds for sale, whether or not species are alien to Taiwan and song attractiveness—as present in all of the possible model combinations. The presence of yellow coloration also had relatively high variable importance (median importance [5th, 95th percentiles] = 0.96 [0.93, 0.98]). The variable importance of the CITES trade status is less than the previous five predictors but higher than the other variables (median importance [5th, 95th percentiles] = 0.71 [0.65, 0.79]). Thus, cheaper birds in Taiwanese pet shops tend to be physically small-bodied, not yellow in colour and not attractive singers, as well as highly available in the local markets, Taiwanese natives, and not included in any CITES Appendices.

The 13 most likely models also included effects of IUCN listing, breeding area and grey colour, although these variables had lower importance than the others in the model ([Table 4](#)). Prices tended to be lower for species less threatened on IUCN status, with larger geographic range sizes and more grey colouration.

Table 1. The ten traded bird species with the largest numbers and ten with highest price recorded for sale in Taiwanese pet shops in our survey.

Family	Species	Number of birds	Price (US\$)	Status
Columbidae	<i>Streptopelia tranquebarica</i>	7420	1.63	Native
Zosteropidae	<i>Zosterops japonicus</i>	6381	35.50	Native
Estrildidae	<i>Lonchura punctulata</i>	2649	1.50	Native
Psittacidae	<i>Melopsittacus undulatus</i>	956	7.70	Alien
Estrildidae	<i>Padda oryzivora</i>	928	10.47	Alien
Sturnidae	<i>Acridotheres grandis</i>	722	10.00	Alien
Sturnidae	<i>Acridotheres fuscus</i>	492	6.83	Alien
Pycnonotidae	<i>Pycnonotus sinensis</i>	465	10.03	Native
Estrildidae	<i>Lonchura striata</i>	437	2.40	Native
Estrildidae	<i>Erythrura gouldiae</i>	405	13.03	Alien
Psittacidae	<i>Guaruba guarouba</i>	5	8000.00	Alien
Psittacidae	<i>Cacatua leadbeateri</i>	2	7500.00	Alien
Psittacidae	<i>Amazona oratrix</i>	1	6666.67	Alien
Psittacidae	<i>Ara macao</i>	1	4000.00	Alien
Psittacidae	<i>Cacatua moluccensis</i>	2	3333.33	Alien
Psittacidae	<i>Ara chloropterus</i>	2	2083.33	Alien
Psittacidae	<i>Cacatua ophthalmica</i>	1	2000.00	Alien
Psittacidae	<i>Primolius auricollis</i>	5	1916.67	Alien
Psittacidae	<i>Ara ararauna</i>	18	1818.17	Alien
Psittacidae	<i>Pionites leucogaster</i>	4	1777.77	Alien

Table with numbers of birds recorded for sale, their price (US\$) and whether species are alien or native to Taiwan (Status).

doi:10.1371/journal.pone.0127482.t001

The analyses of univariate models (Tables 2 and 3), the global PGLS model (median λ [5th, 95th percentiles] = 0.92 [0.90, 0.94] and the 13 most likely PGLS models (Table 4) all identified a high degree of phylogenetic correlation (Pagel's λ) in the models: price tended to be similar for closely related species (Fig 2).

Discussion

The most important correlate of variation in the number of invasive alien species across countries is currently the volume of imported goods [17], suggesting that international trade is now the primary driver of vertebrate species introductions. The bird trade in East and South-East Asia is economically lucrative and active. The invasion pathway for birds has been well-studied in Western countries, but in Eastern countries birds are traded for different reasons, and trade concerns different species with different characteristics traded in different quantities [20,57,58]. It is particularly important to understand the trade in response to Eastern market demands, in comparison to Western trade, which has been curtailed since the unpopularity of past Acclimatisation Society activity and, currently, by wildlife trade bans. This is the primary motivation behind our analyses of the variation in price of bird species for sale in Taiwanese pet shops. The price of birds is a good indicator of the constraints and demands of the pet market, which allows us to identify the characteristics of traded species, and hence a key element in the process of human-mediated bird invasions in East and South-East Asia.

Our analyses revealed that the price of species in pet shops is strongly associated with their availability. Thus, species that are for sale in large numbers, are native to Taiwan, and with healthy wild populations available for international trade (as assessed by CITES) all fetch a relatively low price in Taiwan. The first two of these variables were present in all of the identified

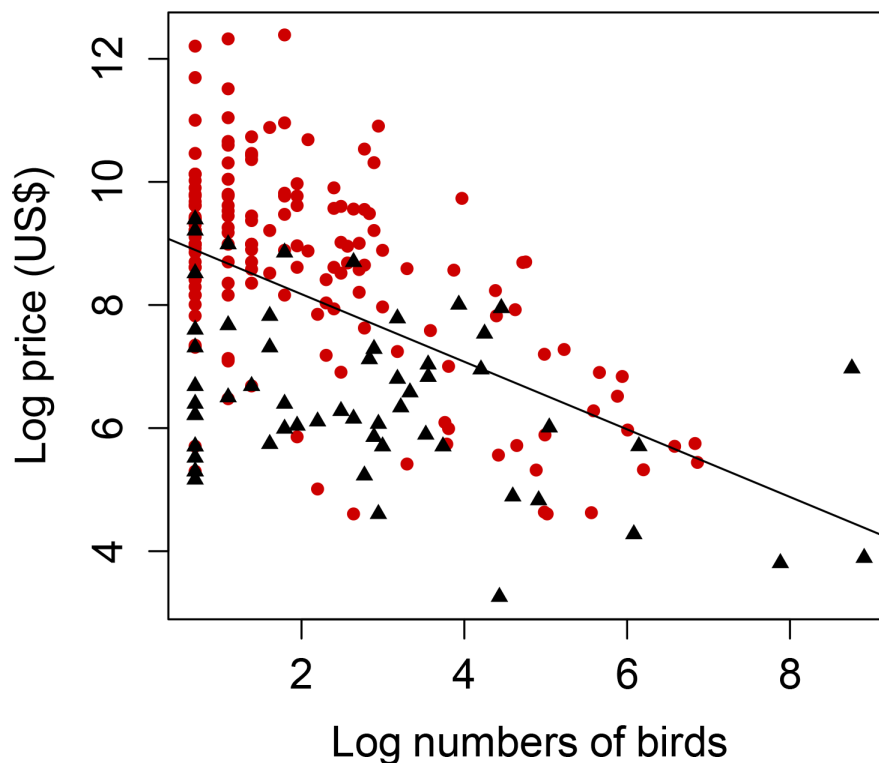


Fig 1. The relationship between log price and log abundance of alien and native species recorded in Taiwan. Red circles indicate the alien species and black triangles indicate the native species for sale in Taiwanese pet shops. The line indicates the best fit relationship between log price (in Taiwanese Dollars, or TWD) and log numbers irrespective of native or alien status ($r^2 = 0.27$, $n = 217$). Both transformations are natural logarithms.

doi:10.1371/journal.pone.0127482.g001

Table 2. The relationship between the log-transformed prices of traded bird species in Taiwanese pet shops and the predictor variables.

Variables	Estimate	t value	Lambda (λ)	p value
Log number for sale	-0.14 [-0.15, -0.14]	-6.91 [-7.01, -6.77]	0.94 [0.94, 0.95]	< 0.001
Log mass	0.93 [0.89, 0.98]	6.27 [5.83, 6.89]	0.93 [0.92, 0.94]	< 0.001
Log geographic range	-0.18 [-0.2, -0.16]	-2.67 [-2.95, -2.44]	0.96 [0.95, 0.97]	0.008 [0.004, 0.01]
Song attractiveness	-0.09 [-0.1, -0.08]	-2.09 [-2.32, -1.91]	0.97 [0.96, 0.97]	0.03 [0.02, 0.05]
Status (native/alien)	-0.54 [-0.55, -0.52]	-2.5 [-2.7, -2.28]	0.95 [0.95, 0.96]	0.01 [0.007, 0.02]
CITES (listed/not)	0.69 [0.66, 0.72]	3.92 [3.71, 4.09]	0.94 [0.93, 0.95]	<0.001
Realm (Australasia, n = 38)	0.06 [0.04, 0.1]	0.27 [0.16, 0.41]	0.95 [0.95, 0.96]	0.78 [0.67, 0.86]
Realm (Indo-Malay, n = 64)	-0.35 [-0.39, -0.31]	-1.42 [-1.57, -1.23]	0.95 [0.95, 0.96]	0.15 [0.11, 0.21]
Realm (Neotropic, n = 41)	0.4 [0.37, 0.43]	1.2 [1.09, 1.29]	0.95 [0.95, 0.96]	0.23 [0.2, 0.27]
Realm (Palearctic, n = 52)	-0.45 [-0.47, -0.42]	-2.16 [-2.34, -1.95]	0.95 [0.95, 0.96]	0.03 [0.02, 0.05]

The averaged coefficients (median estimates [5th, 95th percentiles]) from univariate analysis on the relationship between the log-transformed price of bird species for sale in Taiwanese pet shops ($n = 217$) and the predictor variables in the first column, were calculated over 100 randomly selected likely phylogenetic trees (see [Methods](#) for more details). Parameter Lambda (λ) indicates the strength of phylogenetic signal in the dependent variable (price of birds for sale in Taiwan).

doi:10.1371/journal.pone.0127482.t002

Table 3. The relationship between the price of traded bird species in Taiwanese pet market and the percentage of a given colour found on the species, and colour diversity.

Colour (percentage of colour)	Estimate	t value	p value	Lambda (λ)
Black	3.95 [3.51, 4.34]	2.03 [1.79, 2.27]	0.04 [0.02, 0.07]	0.94 [0.93, 0.95]
Blue	1.45 [1.14, 1.74]	0.93 [0.76, 1.11]	0.34 [0.26, 0.44]	0.94 [0.93, 0.95]
Green	-2.12 [-2.64, -1.53]	-1.61 [-2.06, -1.14]	0.1 [0.04, 0.25]	0.94 [0.94, 0.95]
Red	4.49 [4.15, 4.82]	2.41 [2.24, 2.61]	0.01 [0.009, 0.02]	0.94 [0.93, 0.95]
Orange	-2.3 [-2.54, -1.86]	-1.31 [-1.44, -1.04]	0.19 [0.15, 0.29]	0.94 [0.94, 0.95]
Yellow	4 [3.72, 4.28]	2.89 [2.69, 3.06]	0.004* [0.002, 0.007]	0.95 [0.94, 0.95]
Brown	-2.69 [-3.05, -2.24]	-1.75 [-2.04, -1.46]	0.08 [0.04, 0.14]	0.94 [0.93, 0.95]
Pink	5.25 [4.55, 6.25]	2 [1.66, 2.57]	0.04 [0.01, 0.09]	0.95 [0.94, 0.95]
White	-0.55 [-1.17, 0.16]	-0.22 [-0.47, 0.06]	0.82 [0.63, 0.98]	0.94 [0.93, 0.95]
Grey	-6.2 [-6.64, -5.79]	-3.32 [-3.55, -3.12]	0.001* [0.0004, 0.002]	0.94 [0.93, 0.95]
Ivory	0.14 [-0.44, 0.82]	0.05 [-0.17, 0.33]	0.91 [0.72, 0.98]	0.94 [0.93, 0.95]
Colour Diversity	0.007 [0.002, 0.01]	0.26 [0.07, 0.59]	0.78 [0.55, 0.94]	0.94 [0.04, 0.95]

The averaged coefficients (median estimates, [5th, 95th percentiles]) from univariate analysis on the relationship between the log-transformed price of bird species for sale in Taiwanese pet shops and the arcsine square root transformed percentage of a given colour found on the species, and colour diversity (in the first column), were calculated over 100 likely phylogenetic trees. Parameter Lambda (λ) indicates the strength of phylogenetic signal in the price of birds for sale in Taiwan.

*significant applying a Bonferroni correction for multiple tests; N = 217 in each case.

doi:10.1371/journal.pone.0127482.t003

most likely models, while the third was present in most of these models (Table 4). All also have model-averaged regression coefficients that differ significantly from zero with a high degree of confidence (Tables 2 and 4). For example, the Eurasian Tree Sparrow *P. montanus*, that is a common resident native species in Taiwan [59], is not included in CITES Appendices, and is commonly sold species in Taiwanese pet shops, it costs around US\$0.86 per individual. In contrast, the Golden Parakeet *G. guarouba* is native to South America (i.e. alien), is included on CITES Appendix I (at the time of our survey), and costs around US\$8,000 per individual.

Alien species are probably more expensive because of reduced availability, as there are fewer alien birds for sale in Taiwanese pet shops than natives. Alien species must be imported from overseas [60], and higher prices presumably reflect importation costs (Fig 1). Moreover, higher prices could also reflect the scarcity of captive-bred individuals, particularly for species included on CITES Appendix I. The price inflation for species as alien versus as native is also illustrated by the comparative prices of species in bird markets in different countries. For example, the Red-crested Cardinal *Paroaria coronata* and Saffron Finch *Sicalis flaveola* are both native to South America. Their prices in local markets in Brazil were US\$83.33 and US\$16.11, respectively (where the lowest recorded price of any species is US\$3.33) [61], in comparison to US\$750 and US\$122.23 for the same species in Taiwanese pet markets (where the lowest recorded price of any species is US\$0.86).

Whether or not a species was listed in the CITES appendices (and therefore restricted in international trade) has significant effects on the price of species in the Taiwanese pet market (Tables 2 and 4), albeit that the variable importance for CITES listing is less than that of the five primary predictors of price (Table 4). The sources of species in pet shops (captive-bred or wild-caught) are often unclear [28], but the bulk of pet trade is made up of wild-caught species [16,27,57]. Not only is the international trade of CITES-listed species restricted, but they are also likely to fall under local regulations such as the Wildlife Conservation Act in Taiwan [20,62], and so we would expect an effect of CITES listing on price. Courchamp *et al.* [32] also

Table 4. Model coefficients in PGLS models for determinants of the price of birds in Taiwanese pet shops.

Intercept	Whether listed in CITES appendices	Grey colour	Log breeding range	Log numbers	Log mass	IUCN status	Song attractiveness
2.09 [1.99, 2.2]	0.28 [0.26, 0.31]	−2.61 [−2.9, −2.29]		−0.1 [−1.1, −1.1]	0.61 [0.57, 0.66]		−0.07 [−0.08, −0.06]
1.99 [1.89, 2.1]	0.28 [0.25, 0.31]			−0.1 [−0.1, −0.1]	0.64 [0.59, 0.69]		−0.07 [−0.07, −0.06]
2.1 [2, 2.21]	0.2 [0.251, 0.3]	−2.57 [−2.82, −2.24]		−0.1 [−1.1, −1.1]	0.61 [0.56, 0.66]	0.06 [0.04, 0.08]	−0.07 [−0.08, −0.06]
2.23 [2.1, 2.36]	0.28 [2.56, 0.31]	−2.61 [−2.89, −2.28]	−0.02 [−0.2, −0.008]	−0.1 [−1.1, −1.1]	0.61 [0.56, 0.66]		−0.07 [−0.08, −0.06]
2.02 [1.93, 2.14]		−2.51 [−2.83, −2.19]		−0.11 [−0.11, −0.1]	0.67 [0.62, 0.71]		−0.07 [−0.07, −0.06]
1.99 [1.9, 2.11]	0.27 [0.23, 0.3]			−0.1 [−0.11, −0.1]	0.64 [0.59, 0.69]	0.07 [0.04, 0.08]	−0.06 [−0.07, −0.05]
1.92 [1.82, 2.04]				−0.11 [−0.11, −0.11]	0.69 [0.65, 0.74]		−0.06 [−0.07, −0.05]
2.13 [2.01, 2.25]	0.27 [0.24, 0.3]		−0.02 [−0.03, −0.009]	−0.1 [−0.11, −0.1]	0.64 [0.59, 0.69]		−0.07 [−0.07, −0.06]
2.04 [1.94, 2.15]		−2.46 [−2.77, −2.12]		−0.11 [−0.11, −0.1]	0.66 [0.61, 0.7]	0.07 [0.05, 0.09]	−0.06 [−0.07, −0.05]
1.94 [1.83, 2.06]				−0.11 [−0.11, −0.11]	0.68 [0.63, 0.73]	0.08 [0.05, 0.09]	−0.06 [−0.07, −0.05]
2.16 [2.05, 2.27]	0.29 [0.26, 0.31]	−3.46 [−3.73, −3.14]		−0.1 [−0.1, −0.1]	0.61 [0.56, 0.65]		−0.07 [−0.08, −0.07]
2.15 [2.02, 2.29]	0.28 [0.24, 0.3]	−2.58 [−2.85, −2.24]	−0.009 [−0.01, 0.003]	−0.1 [−0.1, −0.1]	0.61 [0.56, 0.66]	0.05 [0.03, 0.07]	−0.07 [−0.08, −0.06]
2.19 [2.06, 2.32]		−2.51 [−2.82, −2.18]	−0.02 [−0.03, −0.01]	−0.11 [−0.11, −0.1]	0.66 [0.61, 0.7]		−0.07 [−0.07, −0.06]
Importance	0.71 [0.65, 0.79]	0.6 [0.5, 0.7]	0.23 [0.16, 0.29]	1	1	0.29 [0.22, 0.35]	1
Status (native/exotic)	Yellow colour	Model ranks			ΔAIC	Lambda (λ)	Akaike weight
		Rank 1	Rank 2	Rank 3			
−0.24 [−0.27, −0.22]	2.74 [2.56, 2.95]	97	3	0	0 [0, 1]	0.92 [0.9, 0.93]	0.18 [0.154, 0.21]
−0.25 [−0.27, −0.22]	3.17 [0.002, 3.36]	3	84	9	0.71 [0.02, 1.44]	0.92 [0.9, 0.93]	0.12 [0.08, 0.17]
−0.24 [−0.27, −0.21]	2.66 [2.48, 2.85]	0	11	71	1.37 [0.98, 1.88]	0.92 [0.91, 0.94]	0.09 [0.07, 0.11]
−0.24 [−0.27, −0.22]	2.7 [2.52, 2.91]	0	0	2	1.82 [1.43, 2.11]	0.92 [0.9, 0.94]	0.07 [0.06, 0.09]
−0.24 [−0.27, −0.21]	2.74 [2.54, 2.95]	0	2	9	1.86 [1.33, 2.48]	0.92 [0.91, 0.94]	0.07 [0.04, 0.09]
−0.25 [−0.27, −0.23]	3.08 [2.91, 3.24]	0	0	8	2.03 [1.31, 2.66]	0.92 [0.91, 0.93]	0.06 [0.04, 0.08]
−0.25 [−0.27, −0.22]	3.16 [2.98, 3.36]	0	0	1	2.41 [1.49, 3.23]	0.92 [0.91, 0.94]	0.05 [0.03, 0.07]
−0.25 [−0.27, −0.22]	3.13 [2.96, 3.32]	0	0	0	2.54 [1.83, 3.10]	0.92 [0.91, 0.93]	0.05 [0.03, 0.06]
−0.24 [−0.27, −0.21]	2.65 [2.45, 2.85]	0	0	0	2.96 [2.13, 3.92]	0.93 [0.91, 0.94]	0.04 [0.02, 0.05]
−0.24 [−0.27, −0.21]	3.04 [2.87, 3.24]	0	0	0	3.47 [2.43, 4.31]	0.93 [0.91, 0.94]	0.03 [0.02, 0.04]
−0.28 [−0.31, −0.26]		0	0	0	3.48 [2.69, 4.28]	0.91 [0.89, 0.93]	0.03 [0.02, 0.04]

Status (native/ exotic)	Yellow colour	Model ranks			Δ AIC	Lambda (λ)	Akaike weight
		Rank 1	Rank 2	Rank 3			
–0.24 [–0.27, –0.21]	2.65 [2.47, 2.85]	0	0	0	3.51 [3.03, 4.08]	0.92 [0.91, 0.93]	0.03 [0.02, 0.04]
–0.25 [–0.27, –0.21]	2.7 [2.49, 2.91]	0	0	0	3.53 [2.84, 4.42]	0.92 [0.91, 0.94]	0.03 [0.02, 0.04]
1	0.96 [0.93, 0.98]						

Model coefficients (median estimates, [5th, 95th percentiles]) in PGLS models (Δ AIC from the best model is less than 4) for determinants of the price of birds in Taiwanese pet shops, were calculated over 100 randomly selected likely phylogenetic trees. Model ranks (top 3 rankings shown) are based on Δ AIC. Parameter Lambda (λ) indicates the strength of phylogenetic signal in the price of birds for sale in Taiwan for each likely model. IUCN status divided species by whether they are classed in the category of least concern.

doi:10.1371/journal.pone.0127482.t004

showed that CITES listed amphibian and reptile species fetched a higher price in French trade than equivalent non-CITES species.

Larger-bodied species sell for a higher price in the Taiwanese pet market (Tables 2 and 4). Large-bodied species tend to have lower reproductive rates, require more food and space, and live at lower densities in the wild than smaller-bodied species [63]. All of these factors will increase the costs of dealing in these species for pet shop owners, bird breeding farms, and bird catchers and export/importers. Large-bodied species also tend to be long-lived. For example, some of the larger parrot species have lifespans similar to humans [64–66]. Buying such a bird may be a once in a lifetime purchase, raising the price that is likely to be charged. A well-trained parrot may command a very high price, especially if it can imitate human language [42]. This probably explains why species in birds such as macaws (*Ara*), amazons (*Amazona*) and cockatoos (*Cacatua*) fetch a high price in Taiwan (Fig 2).

We predicted a positive effect of the attractiveness of a bird's song on its price, because many species are traded for their ability as songsters and previous analysis showed that species for sale in Taiwanese pet shops had more aesthetically-pleasing songs than expected by chance [67]. In fact, we found the reverse relationship. It is possible that a higher demand for species with attractive songs has led to an abundance of such species in the market, and hence an associated lower price. However, the relationship is more likely to be a consequence of some birds, such as parrots, not being traded for their song, but for other reasons, such as the ability to mimic human voice [42]. If we limit analysis to song birds (species in the order Passeriformes), then the negative relationship between song attractiveness and price disappears (univariate PGLS: estimate median [5th, 95th percentiles] = –0.03 [–0.04, –0.01], $t = -0.48$ [–0.66, 0.22], $p = 0.063$ [0.51, 0.82], $\lambda = 0.90$ [0.88, 0.93]).

We also found some evidence that the price of bird species in Taiwan is influenced by plumage colour, and specifically that species with a higher proportion of yellow colouration fetch a higher price, while species with a higher proportion of grey colouration fetch a lower price (Tables 3 and 4). Cultural factors may be key in driving such colour preferences [68]. For example, the colour grey has associations with low value in Asian countries (e.g., China and Japan), but associations with high quality and expense in the U.S. [69]. In traditional Chinese culture, nature is composed of five elements—wood, fire, earth, metal and water—and each element has a colour and a compass point associated with it [70]. Yellow exemplifies the earth and represents the centre of the compass, and is therefore of high importance. Yellow was the colour for Imperial China: yellow was an exclusive colour of the imperial family during some Dynasties (e.g.,

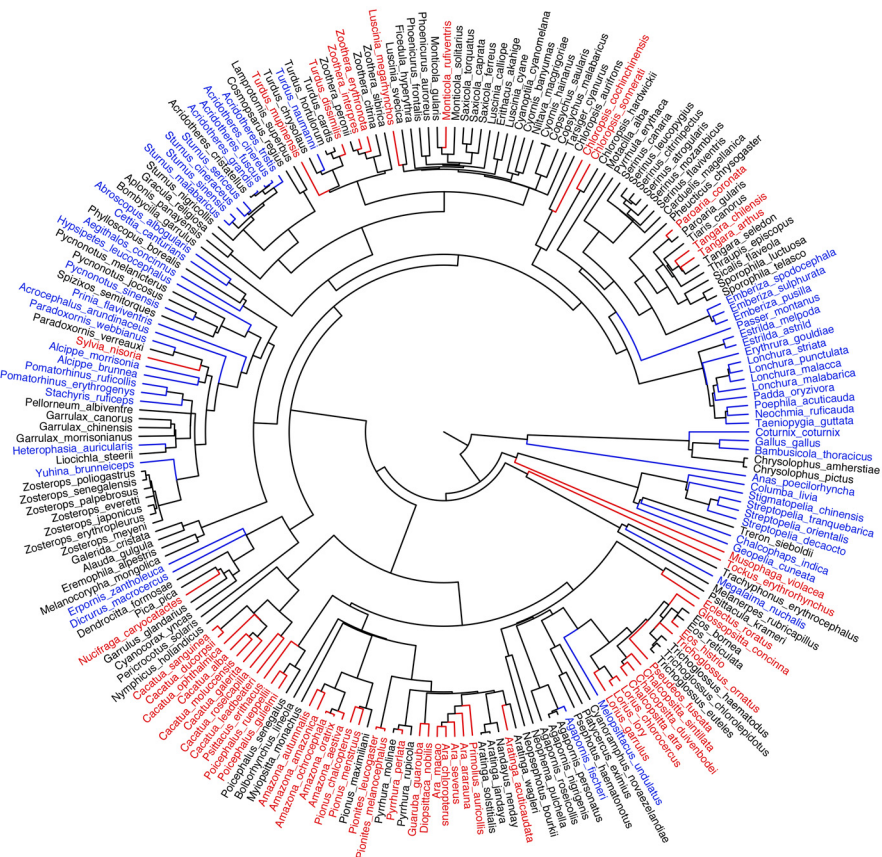


Fig 2. A phylogeny of the bird species included in our analysis. Species were classed in three ranges by the price for every identified species ($n = 217$). Expensive bird species (species in the top quartile for price; $n = 56$) are marked in red; cheap bird species (species in the bottom quartile for price; $n = 55$) are marked in blue. The phylogeny was generated from birdtree.org (Jetz et al. 2012), visualised in FigTree v1.4.0.

doi:10.1371/journal.pone.0127482.g002

the Qing dynasty)[71] and is a colour venerated in Buddhism. Taiwan has a Chinese-influenced culture, and it seems that yellow plumage convinces people to pay a higher price for bird species in the Taiwanese pet trade.

The price of birds was apparently less affected by variables relating to the realm of origin or native breeding range size. The effects of realm of origin and native breeding range size, which we predicted would affect price through availability, are presumably captured more directly by other availability measures, such as number for sale and whether or not a species is native.

Phylogenetic analysis identified a strong phylogenetic correlation in bird price. Related species thus tend to vary in price in a similar manner, which suggests that the factors that determine price are similar for related species. In fact, at least three of the four main predictors of price would be expected to show strong phylogenetic autocorrelation: body mass [72], threat status [73], and alien status, the last given that several of the families with birds for sale in Taiwan have no native species (e.g. parrots, turacos, tanagers). The phylogenetic correlation for analysis of the fourth main predictor—numbers of birds for sale—is also strong (Table 2). However, why this should be so is less clear, especially given that the abundances of bird species in the wild tend to show relatively low phylogenetic correlation (e.g. [74,75]). It may be that some bird taxa are easier to catch, survive better in transportation, and/or are easier to breed in captivity.

Implications

Alien invasive species have a wide range of environmental impacts [76], and are one of the largest threats to the persistence of native species [8]. In Taiwan, for example, the endemic Styan's bulbul *Pycnonotus taivanus* is categorised as vulnerable on the IUCN Red List due to hybridization with the introduced alien Chinese bulbul *P. sinensis*. The aim of this study was to provide insight into the contribution of the cage bird trade to bird invasions. Assuming that cheaper birds are more likely to be liberated (or conversely that more expensive birds are less likely to be so), and species more commonly sold also more likely to escape or be released, then one might conclude from our results that pet shop birds represent a low risk for alien bird invasions in Taiwan. The cheapest species in pet shops are native, as are the species most likely to be bought for prayer release. Prayer birds typically need to be readily and cheaply available [77]. Prayer animals are released regardless of whether or not they are alien to Taiwan, but alien bird species are often priced higher in the shops. Thus, cheaper native species are presumably much easier to source for shops, especially for sale for prayer release. The three most abundant species recorded in pet shops in our survey (Table 1) are all frequently released in religious ceremonies in Taiwan, and are native. All of this suggests that most of the individuals and species likely to be liberated in Taiwan do not present a risk of alien invasion because they are not alien.

Pet shop birds nevertheless still present a non-trivial invasion risk. Most of the species sold are alien (68%), and some of them are available for sale in reasonable numbers (Table 1) and at relatively low price (Figs 1 and 2). There is therefore the potential and opportunity for some alien species to be introduced to the wild, and in the sort of numbers required to promote population establishment (which is a positive function of the number of birds introduced [11,24,25,78]). This likelihood may be further increased by the dynamics of the bird trade. Robinson [79] showed that a high price for rare species can drive oversupply when traders and breeders rush to fulfil demand, with the result that market prices can decline rapidly. When demand is saturated and the market is low on profit, owners of breeding farms or pet shops would be more likely to release their birds (although not while species still sell for higher prices than other similar species). Alternatively, to minimise losses arising from the maintenance of unsold and unsellable birds in shops, pet shops may sell birds at a lower price for religious events.

Previous studies have also shown that when the sale of wild birds is prohibited (i.e. due to avian influenza), vendors were more likely to move to the illegal trade, or to release birds directly into the wild [80]. Many Asian countries have gone through outbreaks of avian flu since 2004 [81], and the bird trade in Taiwan is at risk of being affected by the disease or trade bans of nearby countries. A trade ban could therefore cause species to be released by vendors, whether alien or native. We suspect that species released this way would be likely to be those with lower economic value and to be more abundant in shops. However, maintaining higher value alien species costs more, and so owners may release the birds when their economic situation worsens. Therefore, we would not exclude the possibilities of pricier alien species also being released into the wild in these particular circumstances. We will explore the influence of price on the likelihood of introduction and establishment in a subsequent paper.

Solutions for preventing alien bird releases could, and do, target both the sources of the birds in trade, and the customers sustaining the trade. The numbers of traded individuals could be reduced through control of the sources of birds (wild capture and illegal imports); some such controls are already in place in Taiwan [20,62]. Educating the general population about the risk of releasing alien pet birds may have limited impact in the Asian context, because many purchases are for religious and cultural reasons and not only for companionship or

status. Campaigns are already undertaken to discourage prayer release in Taiwan [77,82], although this practice continues to be commonplace there. An alternative approach (inspired by Venerable Bengkong Shi [83]) would be to co-opt the religious reasons for buying and releasing animals, and so gradually to replace the ceremonial release of randomly chosen (and potentially harmful) prayer animals with planned releases of native species for conservation purposes, such as re-introductions. This process could incorporate ritual ceremonies to pray for the welfare of the animals before they are released.

Supporting Information

S1 Appendix. The species recorded for sale in Taiwanese pet shops (first visit) in our survey, together with the variables used in our analysis.

(CSV)

S2 Appendix. Data on bird coloration used in our analysis.

(CSV)

Acknowledgments

We thank Y. Wang and William Chi for providing information on the wild bird trade in Taiwan, and David Orme and Valerie Olson for supplying data on geographic ranges and body size.

Author Contributions

Conceived and designed the experiments: SS PC TMB. Performed the experiments: SS. Analyzed the data: SS. Contributed reagents/materials/analysis tools: SS MVI PC TMB. Wrote the paper: SS. Revised draft critically for important intellectual content: PC MVI TMB.

References

1. Lockwood JL, Cassey P, Blackburn T. The role of propagule pressure in explaining species invasions. *Trends Ecol Evol.* 2005; 20: 223–228. doi: [10.1016/j.tree.2005.02.004](https://doi.org/10.1016/j.tree.2005.02.004) PMID: [16701373](https://pubmed.ncbi.nlm.nih.gov/16701373/)
2. Blackburn TM, Lockwood JL, Cassey P. *Avian invasions: the ecology and evolution of exotic birds.* Oxford; New York: Oxford University Press; 2009.
3. Davis MA. *Invasion biology.* Oxford; New York: Oxford University Press; 2009.
4. Wilson JRU, Dormontt EE, Prentis PJ, Lowe AJ, Richardson DM. Something in the way you move: dispersal pathways affect invasion success. *Trends Ecol Evol.* 2009; 24: 136–144. doi: [10.1016/j.tree.2008.10.007](https://doi.org/10.1016/j.tree.2008.10.007) PMID: [19178981](https://pubmed.ncbi.nlm.nih.gov/19178981/)
5. Kolar CS, Lodge DM. Progress in invasion biology: predicting invaders. *Trends Ecol Evol.* 2001; 16: 199–204. PMID: [11245943](https://pubmed.ncbi.nlm.nih.gov/11245943/)
6. Blackburn TM. Avian Extinction and Mammalian Introductions on Oceanic Islands. *Science.* 2004; 305: 1955–1958. doi: [10.1126/science.1101617](https://doi.org/10.1126/science.1101617) PMID: [15448269](https://pubmed.ncbi.nlm.nih.gov/15448269/)
7. Frenot Y, Chown SL, Whinam J, Selkirk PM, Convey P, Skotnicki M, et al. Biological invasions in the Antarctic: extent, impacts and implications. *Biol Rev.* 2005; 80: 45–72. doi: [10.1017/S1464793104006542](https://doi.org/10.1017/S1464793104006542) PMID: [15727038](https://pubmed.ncbi.nlm.nih.gov/15727038/)
8. Clavero M, Garciaberthou E. Invasive species are a leading cause of animal extinctions. *Trends Ecol Evol.* 2005; 20: 110–110. doi: [10.1016/j.tree.2005.01.003](https://doi.org/10.1016/j.tree.2005.01.003) PMID: [16701353](https://pubmed.ncbi.nlm.nih.gov/16701353/)
9. Lockwood JL, McKinney ML, editors. *Biotic homogenization.* New York: Kluwer Academic/Plenum Publishers; 2001.
10. Bush ER, Baker SE, Macdonald DW. Global Trade in Exotic Pets 2006–2012: Exotic Pet Trade. *Conserv Biol.* 2014; 28: 663–676. doi: [10.1111/cobi.12240](https://doi.org/10.1111/cobi.12240) PMID: [24661260](https://pubmed.ncbi.nlm.nih.gov/24661260/)
11. Blackburn TM, Pyšek P, Bacher S, Carlton JT, Duncan RP, Jarošík V, et al. A proposed unified framework for biological invasions. *Trends Ecol Evol.* 2011; 26: 333–339. doi: [10.1016/j.tree.2011.03.023](https://doi.org/10.1016/j.tree.2011.03.023) PMID: [21601306](https://pubmed.ncbi.nlm.nih.gov/21601306/)

12. Meyerson LA, Mooney HA. Invasive alien species in an era of globalization. *Front Ecol Environ*. 2007; 5: 199–208. doi: [10.1890/1540-9295\(2007\)5\[199:IASIAE\]2.0.CO;2](https://doi.org/10.1890/1540-9295(2007)5[199:IASIAE]2.0.CO;2)
13. Hulme PE. Trade, transport and trouble: managing invasive species pathways in an era of globalization. *J Appl Ecol*. 2009; 46: 10–18. doi: [10.1111/j.1365-2664.2008.01600.x](https://doi.org/10.1111/j.1365-2664.2008.01600.x)
14. García-Díaz P, Ross JV, Ayres C, Cassey P. Understanding the biological invasion risk posed by the global wildlife trade: propagule pressure drives the introduction and establishment of Nearctic turtles. *Glob Change Biol*. 2014; n/a–n/a. doi: [10.1111/gcb.12790](https://doi.org/10.1111/gcb.12790)
15. Lee TW, Shieh BS. Pet sales of exotic estrildid birds in relation to the field-records in Taiwan. *Endem Species Res*. 2005; 7: 1–12.
16. Carrete M, Tella J. Wild-bird trade and exotic invasions: a new link of conservation concern? *Front Ecol Environ*. 2008; 6: 207–211. doi: [10.1890/070075](https://doi.org/10.1890/070075)
17. Westphal MI, Browne M, MacKinnon K, Noble I. The link between international trade and the global distribution of invasive alien species. *Biol Invasions*. 2008; 10: 391–398. doi: [10.1007/s10530-007-9138-5](https://doi.org/10.1007/s10530-007-9138-5)
18. Colunga-Garcia M, Haack R, Magarey R, Borchert D. Understanding trade pathways to target biosecurity surveillance. *NeoBiota*. 2013; 18: 103–118. doi: [10.3897/neobiota.18.4019](https://doi.org/10.3897/neobiota.18.4019)
19. International Traffic Network. Perceptions, conservation, and management of wild birds in trade. Cambridge, U.K: TRAFFIC International; 1992.
20. Severinghaus LL, Chi L. Prayer animal release in Taiwan. *Biol Conserv*. 1999; 89: 301–304.
21. Agoramoorthy G, Hsu MJ. Ritual Releasing of Wild Animals Threatens Island Ecology. *Hum Ecol*. 2007; 35: 251–254.
22. Environment and Animal Society of Taiwan (EAST), Kaohsiung Teacher's Association. The practices of prayer animal releases conducted by religious organisations in Taiwan. [Internet]. Taiwan: Environment and Animal Society of Taiwan; 2004. Available: http://www.east.org.tw/that_content.php?s_id=14&m_id=0&id=89
23. Blackburn TM, Duncan RP. Establishment patterns of exotic birds are constrained by non-random patterns in introduction. *J Biogeogr*. 2001; 28: 927–939.
24. Cassey P, Blackburn TM, Russell GJ, Jones KE, Lockwood JL. Influences on the transport and establishment of exotic bird species: an analysis of the parrots (Psittaciformes) of the world. *Glob Change Biol*. 2004; 10: 417–426. doi: [10.1111/j.1529-8817.2003.00748.x](https://doi.org/10.1111/j.1529-8817.2003.00748.x)
25. Lockwood JL, Cassey P, Blackburn T. The role of propagule pressure in explaining species invasions. *Trends Ecol Evol*. 2005; 20: 223–228. doi: [10.1016/j.tree.2005.02.004](https://doi.org/10.1016/j.tree.2005.02.004) PMID: [16701373](https://pubmed.ncbi.nlm.nih.gov/16701373/)
26. Simberloff D. The Role of Propagule Pressure in Biological Invasions. *Annu Rev Ecol Evol Syst*. 2009; 40: 81–102. doi: [10.1146/annurev.ecolsys.110308.120304](https://doi.org/10.1146/annurev.ecolsys.110308.120304)
27. Nash SV. Sold for a Song: The Trade in Southeast Asian Non-CITES birds. Cambridge, UK: Traffic International; 1993.
28. Shepherd CR, Stengel CJ, Nijman V. The Export and Re-export of CITES-listed Birds from the Solomon Islands. Southeast Asia, Petaling Jaya, Selangor, Malaysia.: TRAFFIC; 2012.
29. Severinghaus LL. The breeding ecology and dispersal of Taiwan and Light-vented Bulbuls in Kenting National Park. *J Natl Park*. 2007; 17: 27–41.
30. Li S-H, Yeung CK-L, Han L, Le MH, Wang C, Ding P, et al. Genetic introgression between an introduced babbler, the Chinese hwamei *Leucodioptron c. canorum*, and the endemic Taiwan hwamei *L. taewanus*: a multiple marker systems analysis. *J Avian Biol*. 2010; 41: 64–73. doi: [10.1111/j.1600-048X.2009.04719.x](https://doi.org/10.1111/j.1600-048X.2009.04719.x)
31. Yao C-T. The elimination and caged raise of exotic Red-billed Blue Magpie *urocissa erythrorhyncha* in Wu-ling National Forest Recreation Area. Wu-ling National Forest Recreation Area, Taiwan: Endemic Species Research Institute, Council of Agriculture, Executive Yuan, Taiwan; 2007. Report No.: 96-03-8-04.
32. Courchamp F, Angulo E, Rivalan P, Hall RJ, Signoret L, Bull L, et al. Rarity Value and Species Extinction: The Anthropogenic Allee Effect. *PLoS Biol*. 2006; 4: e415. doi: [10.1371/journal.pbio.0040415](https://doi.org/10.1371/journal.pbio.0040415) PMID: [17132047](https://pubmed.ncbi.nlm.nih.gov/17132047/)
33. Blackburn TM, Cassey P, Gaston KJ. Variations on a theme: sources of heterogeneity in the form of the interspecific relationship between abundance and distribution. *J Anim Ecol*. 2006; 75: 1426–1439. doi: [10.1111/j.1365-2656.2006.01167.x](https://doi.org/10.1111/j.1365-2656.2006.01167.x) PMID: [17032375](https://pubmed.ncbi.nlm.nih.gov/17032375/)
34. Gaston KJ, Blackburn TM. Pattern and process in macroecology. Oxford, OX; Malden, MA, USA: Blackwell Science; 2000.
35. Kneiberting E, Van Der Windt HJ, Schoot Uiterkamp AJM. Public decisions on animal species: does body size matter? *Environ Conserv*. 2010; 38: 28–36. doi: [10.1017/S0376892910000755](https://doi.org/10.1017/S0376892910000755)

36. Kellert S. American Attitudes Toward and Knowledge of Animals: An Update. In: Fox M, Mickley L, editors. *Advances in Animal Welfare Science* 1984. Springer Netherlands; 1985. pp. 177–213. Available: http://dx.doi.org/10.1007/978-94-009-4998-0_11
37. Blackburn TM, Gaston KJ. Body Size and Density: The Limits to Biomass and Energy Use. *Oikos*. 1994; 69: 336–339. doi: [10.2307/3546157](https://doi.org/10.2307/3546157)
38. Blackburn TM, Gaston KJ. Special Paper: What Determines the Probability of Discovering a Species?: A Study of South American Oscine Passerine Birds. *J Biogeogr*. 1995; 22: 7. doi: [10.2307/2846069](https://doi.org/10.2307/2846069)
39. Gaston KJ. Rarity [Internet]. Dordrecht: Springer Netherlands; 1994. Available: <http://dx.doi.org/10.1007/978-94-011-0701-3>
40. Frynta D, Lišková S, Bültmann S, Burda H. Being Attractive Brings Advantages: The Case of Parrot Species in Captivity. *PLoS ONE*. 2010; 5: e12568. doi: [10.1371/journal.pone.0012568](https://doi.org/10.1371/journal.pone.0012568) PMID: [20830206](https://pubmed.ncbi.nlm.nih.gov/20830206/)
41. Lišková S, Frynta D. What Determines Bird Beauty in Human Eyes? *Anthrozoos Multidiscip J Interact People Anim*. 2013; 26: 27–41. doi: [10.2752/175303713X13534238631399](https://doi.org/10.2752/175303713X13534238631399)
42. Tella JL, Hiraldo F. Illegal and Legal Parrot Trade Shows a Long-Term, Cross-Cultural Preference for the Most Attractive Species Increasing Their Risk of Extinction. Roberts DL, editor. *PLoS ONE*. 2014; 9: e107546. doi: [10.1371/journal.pone.0107546](https://doi.org/10.1371/journal.pone.0107546) PMID: [25225808](https://pubmed.ncbi.nlm.nih.gov/25225808/)
43. Jetz W, Thomas GH, Joy JB, Hartmann K, Mooers AO. The global diversity of birds in space and time. *Nature*. 2012; 491: 444–448. doi: [10.1038/nature11631](https://doi.org/10.1038/nature11631) PMID: [23123857](https://pubmed.ncbi.nlm.nih.gov/23123857/)
44. Ding T-S, Juan C-S, Lin R-S, Pan C-Y, Tsai Y-J, Wu J, et al. The 2012 CWBF Checklist of the Birds of Taiwan. Taiwan: Bird Record Committee, Chinese Wild Bird Federation; 2012.
45. Gaston KJ. How Large Is a Species' Geographic Range? *Wiley Behalf Nord Soc Oikos*. 1991; 61: 434–438.
46. Gaston KJ. Measuring Geographic Range Sizes. *Wiley Behalf Nord Soc Oikos*. 1994; 17: 198–205.
47. Orme CDL, Davies RG, Olson VA, Thomas GH, Ding TS, Rasmussen PC, et al. Global patterns of geographic range size in birds. *PLoS Biol*. 2006; 4: e208. PMID: [16774453](https://pubmed.ncbi.nlm.nih.gov/16774453/)
48. Blackburn TM, Cassey P. Patterns of non-randomness in the exotic avifauna of Florida. *Divers Distrib*. 2007; 13: 519–526.
49. Olson DM, Dinerstein E, Wikramanayake ED, Burgess ND, Powell GVN, Underwood EC, et al. Terrestrial Ecoregions of the World: A New Map of Life on Earth. *BioScience*. 2001; 51: 933–938. doi: [10.1641/0006-3568\(2001\)051\[0933:TEOTWA\]2.0.CO;2](https://doi.org/10.1641/0006-3568(2001)051[0933:TEOTWA]2.0.CO;2)
50. Dunning JB. *CRC handbook of avian body masses*. Boca Raton, Florida: CRC Press; 1992.
51. Olson VA, Davies RG, Orme CDL, Thomas GH, Meiri S, Blackburn TM, et al. Global biogeography and ecology of body size in birds. *Ecol Lett*. 2009; 12: 249–259. doi: [10.1111/j.1461-0248.2009.01281.x](https://doi.org/10.1111/j.1461-0248.2009.01281.x) PMID: [19245587](https://pubmed.ncbi.nlm.nih.gov/19245587/)
52. IUCN. IUCN Red List of Threatened Species. Version 2013.1. <www.iucnredlist.org> [Internet]. 2013. Available: <www.iucnredlist.org>
53. Blackburn TM, Su S, Cassey P. A Potential Metric of the Attractiveness of Bird Song to Humans. Hauber M, editor. *Ethology*. 2014; 120: 305–312. doi: [10.1111/eth.12211](https://doi.org/10.1111/eth.12211)
54. R Core Team. R: A language and environment for statistical computing [Internet]. Vienna, Austria; 2014. Available: <http://www.R-project.org/>
55. Revell LJ. Phylogenetic signal and linear regression on species data: Phylogenetic regression. *Methods Ecol Evol*. 2010; 1: 319–329. doi: [10.1111/j.2041-210X.2010.00044.x](https://doi.org/10.1111/j.2041-210X.2010.00044.x)
56. Fritz SA, Purvis A. Selectivity in Mammalian Extinction Risk and Threat Types: a New Measure of Phylogenetic Signal Strength in Binary Traits: Selectivity in Extinction Risk. *Conserv Biol*. 2010; 24: 1042–1051. doi: [10.1111/j.1523-1739.2010.01455.x](https://doi.org/10.1111/j.1523-1739.2010.01455.x) PMID: [20184650](https://pubmed.ncbi.nlm.nih.gov/20184650/)
57. Edmunds K, Robertson SI, Few R, Mahood S, Bui PL, Hunter PR, et al. Investigating Vietnam's Ornamental Bird Trade: Implications for Transmission of Zoonoses. *EcoHealth*. 2011; 8: 63–75. doi: [10.1007/s10393-011-0691-0](https://doi.org/10.1007/s10393-011-0691-0) PMID: [21809163](https://pubmed.ncbi.nlm.nih.gov/21809163/)
58. Jepson P, Ladle RJ. Bird-keeping in Indonesia: conservation impacts and the potential for substitution-based conservation responses. *Oryx*. 2005; 39: 442–448.
59. Zhang S, Zheng G, Xu J. Habitat use of urban Tree Sparrows in the process of urbanization: Beijing as a case study. *Front Biol China*. 2008; 3: 308–314. doi: [10.1007/s11515-008-0042-x](https://doi.org/10.1007/s11515-008-0042-x)
60. Low BW. The global trade in native Australian parrots through Singapore between 2005 and 2011: a summary of trends and dynamics. *Emu*. 2014; 114: 277–282.
61. Regueira RFS, Bernard E. Wildlife sinks: Quantifying the impact of illegal bird trade in street markets in Brazil. *Biol Conserv*. 2012; 149: 16–22. doi: [10.1016/j.biocon.2012.02.009](https://doi.org/10.1016/j.biocon.2012.02.009)

62. Agoramoorthy G. Enforcement Challenges of Taiwan's Wildlife Conservation and Animal Protection Laws. *J Int Wildl Law Policy*. 2009; 12: 190–209. doi: [10.1080/13880290903202690](https://doi.org/10.1080/13880290903202690)
63. Bennett PM. *Evolutionary ecology of birds: life histories, mating systems, and extinction*. Oxford ; New York: Oxford University Press; 2002.
64. Wilson LH. Considerations in selecting an appropriate pet bird. *J Am Vet Med Assoc*. 1998; 212: 1222–1225. PMID: [9569156](https://pubmed.ncbi.nlm.nih.gov/9569156/)
65. Robinson JM. The dynamics of avicultural markets. *Environ Conserv*. 2001; 28: 76–85. doi: [10.1017/S0376892901000078](https://doi.org/10.1017/S0376892901000078)
66. Young AM, Hobson EA, Lackey LB, Wright TF. Survival on the ark: life-history trends in captive parrots: Life-history trends in captive parrots. *Anim Conserv*. 2012; 15: 28–43. doi: [10.1111/j.1469-1795.2011.00477.x](https://doi.org/10.1111/j.1469-1795.2011.00477.x) PMID: [22389582](https://pubmed.ncbi.nlm.nih.gov/22389582/)
67. Su S, Cassey P, Blackburn TM. Patterns of non-randomness in the composition and characteristics of the Taiwanese bird trade. *Biol Invasions*. 2014; doi: [10.1007/s10530-014-0686-1](https://doi.org/10.1007/s10530-014-0686-1)
68. Soma I, Saito M. Cross-Cultural Survey on Color Preferences in Three Asian Cities. In: Wapner S, Demick J, Yamamoto T, Takahashi T, editors. *Handbook of Japan-United States Environment-Behavior Research*. Springer US; 1997. pp. 101–112. Available: http://dx.doi.org/10.1007/978-1-4899-0286-3_8
69. Jacobs L, Keown C, Worthley R, Ghymn K-I. Cross-cultural Colour Comparisons: Global Marketers Beware! *Int Mark Rev*. 1991; 8: 21–30. doi: [10.1108/02651339110137279](https://doi.org/10.1108/02651339110137279)
70. Chen G-M. The impact of feng shui on Chinese communication. *China Media Res*. 2007; 3: 102–109.
71. Garrett VM. *Chinese dress: from the Qing dynasty to the present*. Tokyo ; North Clarendon, VT: Tuttle Pub; 2008.
72. Freckleton RP, Harvey PH, Pagel M. Phylogenetic Analysis and Comparative Data: A Test and Review of Evidence. *Am Nat*. 2002; 160: 712–726. doi: [10.1086/343873](https://doi.org/10.1086/343873) PMID: [18707460](https://pubmed.ncbi.nlm.nih.gov/18707460/)
73. Bennett PM, Owens IPF. Variation in Extinction Risk among Birds: Chance or Evolutionary Predisposition? *Proc Biol Sci*. 1997; 264: 401–408. doi: [10.2307/50429](https://doi.org/10.2307/50429)
74. Boehning-Gaese K, Oberrath R. Phylogenetic effects on morphological, life-history, behavioural and ecological traits of birds. *Evol Ecol Res*. 1999; 1: 347–364.
75. Webb TJ, Noble D, Freckleton RP. Abundance-occupancy dynamics in a human dominated environment: linking interspecific and intraspecific trends in British farmland and woodland birds. *J Anim Ecol*. 2007; 76: 123–134. doi: [10.1111/j.1365-2656.2006.01192.x](https://doi.org/10.1111/j.1365-2656.2006.01192.x) PMID: [17184360](https://pubmed.ncbi.nlm.nih.gov/17184360/)
76. Blackburn TM, Essl F, Evans T, Hulme PE, Jeschke JM, Kühn I, et al. A Unified Classification of Alien Species Based on the Magnitude of their Environmental Impacts. *PLoS Biol*. 2014; 12: e1001850. doi: [10.1371/journal.pbio.1001850](https://doi.org/10.1371/journal.pbio.1001850) PMID: [24802715](https://pubmed.ncbi.nlm.nih.gov/24802715/)
77. Environment and Animal Society of Taiwan. The investigation of prayer animal releases in Taiwan [Internet]. Taiwan: Environment & Animal Society of Taiwan and Human Society International; 2009. Available: http://www.east.org.tw/that_content.php?s_id=14&m_id=0&id=89
78. Blackburn TM, Lockwood JL, Cassey P. Avian invasions: the ecology and evolution of exotic birds [Internet]. 2009 [cited 10 Jul 2012]. Available: <http://public.eblib.com/EBLPublic/PublicView.do?ptilID=679349>
79. Robinson JM. The dynamics of avicultural markets. *Environ Conserv*. 2002; 28. doi: [10.1017/S0376892901000078](https://doi.org/10.1017/S0376892901000078)
80. Brooks-Moizer F, Robertson SI, Edmunds K, Bell D. Avian Influenza H5N1 and the Wild Bird Trade in Hanoi, Vietnam. *Ecol Soc*. 2008; 14: 28.
81. Forster P. Ten Years on: Generating Innovative Responses to Avian Influenza. *EcoHealth*. 2013; doi: [10.1007/s10393-013-0887-6](https://doi.org/10.1007/s10393-013-0887-6)
82. Environment and Animal Society of Taiwan. A report of prayer animals in Taiwanese pet shops. Environment and Animal Society of Taiwan; 2004.
83. Nuwer R. A Buddhist Ritual Gets an Ecologically Correct Update. *Audubon*. 2014 January-February. Available: <http://www.audubon.org/magazine/january-february-2014/a-buddhist-ritual-gets-ecologically>

Copyright of PLoS ONE is the property of Public Library of Science and its content may not be copied or emailed to multiple sites or posted to a listserv without the copyright holder's express written permission. However, users may print, download, or email articles for individual use.